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A von Thunen model and case of reed  
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**Bioenergy crop production and climate policies:  
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**Abstract**

We examine bioenergy crop cultivation in a von Thunen framework with homogenous agricultural land and increasing transportation costs by distance. Bioenergy crops can offset emissions from fossil fuels but their cultivation causes nutrient runoff. Increasing transportation costs imply that the fertilizer intensity differs across locations; it also defines the socially and privately optimal extensive margin of bioenergy crop production. If climate benefits are only partially capitalized in the bioenergy crop price, the privately optimal fertilizer application is suboptimal calling for location specific input or output subsidies. A theoretical model is applied to reed canary grass (*Phalaris arundinacea* L.) cultivation in Finland. Reed canary grass offsets emissions from peat in electricity production. Using oats as an alternative crop and permit price € 20/tonne of CO<sub>2</sub> emissions as a proxy for the climate benefits, cultivation of reed canary grass is socially optimal at a distance over 100 kilometres from the power plant and still offsets more than 6.5 tonnes per ha of CO<sub>2</sub> emissions from peat.

**Keywords:** von Thunen model, climate benefits, bioenergy crops, emission permit market, nutrient runoff

**JEL classification:** Q18, Q24, Q25

## 1. Introduction

The agricultural sector is both a source and a sink of greenhouse gases (GHG) (carbon dioxide CO<sub>2</sub>, methane CH<sub>4</sub>, and nitrous oxide N<sub>2</sub>O). The share of agriculture in OECD total national gross emissions in CO<sub>2</sub> equivalents is below 10%, but for methane and nitrous oxide agriculture contributes a major share, 40% and 60%, respectively (OECD 2001). The agricultural sector has the potential for absorbing CO<sub>2</sub> emissions through changes in tillage practices, such as adopting no-till, or changes in land use forms, such as conversion of arable land to grassland.

What is more, agriculture has the potential to offset GHG emissions through production of bioenergy from agricultural biomass (feedstocks). Agricultural biomass can be used to produce energy (heat and electricity), and biofuels, for instance, bioethanol and biogas. The recently established European Union Emission Trading Scheme (EU-ETS) has opened new possibilities for the production of bioenergy crops in Europe. Emissions generated using fossil fuels in energy production belong to the EU-ETS, while emissions from renewable energy do not. Because power plants must pay a permit price for each unit of CO<sub>2</sub> emitted from fossil fuels, the relative energy prices have changed in favour of renewable energy sources and their use is increasing.<sup>2</sup>

For farmers bioenergy crop cultivation provides an interesting new production alternative and for society bioenergy crops provide potential climate benefits and some other positive environmental effects as well. Given that most bioenergy crops are perennial crops with less tillage and lower fertilizer intensity than in cereals production, nutrient runoff to watercourses could be reduced as well. These positive effects are, however, to some extent offset by the fact that transportation of bioenergy crops causes emissions. Moreover, transportation is costly relative to the value of transported feedstock, which implies that the demand for bioenergy crops and the resulting production provides a local

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<sup>2</sup> In the Finnish energy sector peat is typically replaced by bioenergy from agriculture or from forestry. Peat is subject to EU-ETS and its profitability has decreased quite dramatically. Many power plants currently use a mixture of peat and reed canary grass. Some power plants use a mixture of wood chips and reed canary grass to hedge against volatile wood chips prices.

solution at best. Hence, an interesting research question is whether the production of bioenergy crops is socially desirable when environmental and economic effects are taken into account – and to what extent?

In this paper we examine the socially and privately optimal production of bioenergy crops. We develop a von Thunen framework, which is well-designed to analyse economic decisions when distance matters. In our model a power plant, locating at the centre of an area, demands bioenergy crops. The arable land around the power plant is homogenous in terms of its productivity. Transportation is costly and effectively reduces profitability of bioenergy crop production when the distance to the power plant increases. Our analysis comprises both CO<sub>2</sub> emissions to air and nutrient runoff to waterways. We define the optimal fertilizer intensity in each location. For the given alternative crop, we determine the extensive margin of production; this gives the area that is optimal to allocate to the production of bioenergy crops. Finally, we examine the private incentives of bioenergy crop cultivation and design the first-best instruments to encourage bioenergy cultivation to the socially optimal level.

To assess empirically the private profitability and social returns of bioenergy crop production, we apply our von Thunen model to Finnish agriculture. We consider cultivation of reed canary grass (*Phalaris arundinacea* L.), which is regarded as the most suitable bioenergy crop for the climatic conditions in northern Scandinavia (Landström et al. 1996). According to Finnish experiments, reed canary grass can be cultivated on almost all soil types (textural classes) but highest yields are obtained on organic soils. In the Finnish climatic and soil conditions reed canary grass produces 6-8 tons of dry mass per hectare for a time period of 10-12 years (Pahkala et al. 2005). The average net calorific heat value is 2.45-3.94 MWh per ton of dry mass (with moisture content of 15-20%), yielding a mean energy production of 19.2 MWh per ha at harvest of 6 tonnes per ha.

Our work relates to the previous literature as follows. Earlier studies have mainly been empirical and focused on the potential supply of bioenergy crops. Downing and Graham

(1996) and Walsh (1998; 2000) outline an approach to determine the supply function of bioenergy crops subject to relative prices and government subsidies. Walsh et al. (2003) estimate the bioenergy potential of U.S. agriculture and the required prices of bioenergy to compete on land use with other crops. Larsson (2005) applies a supply function approach to Swedish agriculture. While the above studies neglect the climate effects of bioenergy crops, Faaij et al. (1998) include climate damage when assessing externalities of electricity production from biomass and coal. Börjesson (1999a; 1999b) provide a comprehensive representation of the environmental benefits and economic aspects of energy crop cultivation. Unlike the present paper, these studies do not describe the cultivation of bioenergy crops in detail, nor do they provide a theoretical treatment of the subject.

The rest of the paper is organized as follows. In section 2 we develop the theoretical representation of von Thunen model. The empirical model is developed in section 3 and the results of the analysis are presented in section 4. Conclusions and policy implications are given in section 5.

## **2. Bioenergy production in the von Thunen model**

In this section we develop our theoretical model. The model consists of three basic components: private profits from bioenergy production, nutrient runoff damage and CO<sub>2</sub> reduction benefits.

### **2.1 The basic framework**

Consider a local power plant that buys a bioenergy crop from the neighbouring fields and uses it to replace some fossil fuel in electricity production. The power plant's ability to pay for the bioenergy crop depends among other things on the energy content of the bioenergy, substitution possibilities between bioenergy and fossil fuel, as well as on the price of emission permits. The power plant pays a gate price  $\hat{p}$  for bioenergy. We assume that this price is determined competitively. Moreover, we assume that the net

price after transportation cost,  $p$ , decreases in distance,  $k$ , and is given by  $p = \hat{p} - \eta(k)$ , where  $\eta(k)$  is the (convex) transportation cost with  $\eta'(k) > 0$  and  $\eta''(k) > 0$ .

An important question concerning the gate price is whether the climate benefits are actually capitalized in the gate price, or not. Referring to empirical evidence, in Finland the listed gate prices do not include climate benefits but the power plants add to the gate price an emission trading premium, which implies capitalization. However, as we will find in the empirical section, this extra payment does not fully capitalize the climate benefits into the price paid to farmers. We solve the social optimum using the gate price and accounting fully for the social climate benefits. When comparing the social optimum to the private optimum, we account for the emission trading premium and distinguish between partial and full capitalization.

In accordance with the empirical evidence, we assume that the bioenergy crop (such as reed canary grass, or switch grass) is a perennial crop having  $n$ -period long rotation age. Producing the bioenergy crop requires an initial investment in the first year. It includes primary tillage (ploughing), seedbed tillage (harrowing), planting, and crop protection. After the first year, an annual fertilizer input,  $l$ , is applied and the crop provides a positive harvest over  $n-1$  years. The energy crop grows according to a concave response function,  $Q = f(l)$ . We denote the first year's establishment cost by  $\bar{C}$  and the real interest rate by  $r$ . The private profits of a farmer producing bioenergy crop in the relevant locations,  $k = 1, \dots, m$ , of the neighbourhood of the power plant is given by,

$$\pi = \sum_{t=2}^n [(1+r)^{-(t-1)} [pf(l_t) - cl_t]] - \bar{C}. \quad (1)$$

Fertilizer application causes nutrient runoff,  $z$ , to waterways. It depends on the fertilizer application, so that in each period  $t = 2, \dots, n$  we have  $z_t = g(l_t)$  as the per parcel (location) runoff. We regard the first year's fertilizer application as technologically fixed

( $\bar{z}_1 = g(\bar{l}_1)$ ). Let the periodic damage from runoff be  $d$ . The present value of nutrient damages over the whole rotation period,  $D$ , can be expressed as,

$$D = d(g(\bar{l}_1)) + \sum_{t=2}^n (1+r)^{-(t-1)} d(g(l_t)). \quad (2)$$

We assume that the damage is increasing and convex ( $D' > 0$  and  $D'' > 0$ ).

The bioenergy crop can be used to provide a certain amount of electricity. Let a coefficient  $\alpha$  ( $0 < \alpha < 1$ ) denote how much electricity is produced from one metric ton of bioenergy. Furthermore, let  $\hat{\varepsilon}$  ( $0 < \hat{\varepsilon} < 1$ ) denote how much electricity provided by this ton offsets the fossil fuel-based CO<sub>2</sub> emission in electricity production. To ease notation we let  $\varepsilon = \hat{\varepsilon}\alpha$  to denote the offset of CO<sub>2</sub> emissions. Let  $b$  denote the periodic social benefits from reduced CO<sub>2</sub> emissions. These benefits accrue from the second year onwards, and we denote their present value over the rotation period by  $B$

$$B = \sum_{t=2}^n b(\varepsilon f(l_t))(1+r)^{-(t-1)}. \quad (3)$$

The climate benefit function is concave in the CO<sub>2</sub> offsets ( $B' > 0$  and  $B'' < 0$ ).

## 2.2 The socially and privately optimal bioenergy crop production

We start with the socially optimal production and assume that society maximizes the sum of the consumers' and producers' surplus. As we treat the price of bioenergy crop as constant, this amounts to maximizing the sum of equations (1) – (3). Thus, the social welfare function is,

$$SW = \sum_{t=2}^n \left[ (1+r)^{-(t-1)} [pf(l_t) - cl_t - d(g(l_t)) + b(\varepsilon f(l_t))] \right] - \bar{C} - d(g(\bar{l}_1)) \quad (4)$$

Recall, the establishment of bioenergy crop production is technologically fixed. Thus, the key economic problem of the social planner is to choose fertilizer application over periods  $t = 2, \dots, n$  and locations  $k = 1, \dots, m$  so as to maximize (4). The first and the second order conditions of the social optimum can be expressed as,

$$SW_l = pf'(l) - c - d'(\cdot)g_l + b'(\cdot)\epsilon f'(l) = 0 \quad (5)$$

$$SW_{ll} = pf''(l) - d''(\cdot)(g_l)^2 - d'(\cdot)g_{ll} + b''(\cdot)(\epsilon f'(l))^2 + b'(\cdot)\epsilon f''(l) < 0. \quad (6)$$

Note first, that equations (5) and (6) hold for any period  $t = 2, \dots, n$ .<sup>3</sup> Thus, the real interest rate does not affect fertilizer intensity, only the present value of the net returns. Therefore, fertilizer intensity remains the same over the rotation period in each location. Hence, the economic interpretation of equation (5) is conventional. Society chooses the level of fertilizer application by equalizing the social marginal benefits to the social marginal costs, that is,  $pf'(l) + b'(\cdot)\epsilon f'(l) = c + d'(\cdot)g_l$ .<sup>4</sup> Importantly, fertilizer intensity differs between locations because the net price of the energy crop is decreasing in distance  $k$  due to transportation costs, as shown by equation (7):

$$\frac{dl}{dk} = \frac{\eta'(k)f'(l)}{SW_{ll}} < 0. \quad (7)$$

Assuming all external effects are absent, equation (5) reduces to the privately optimal condition,  $p^* f'(l) = c$ , where  $p^*$  refers to the possibility that the price partially or fully includes the offset benefits via emission trading. Naturally, the privately optimal condition requires that the value of the marginal product of fertilizer use equals the input price. Similarly, fertilizer intensity differs between locations.

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<sup>3</sup> Given that fertilizer intensity is the same over the whole period, we hereafter drop the time subscripts  $t$  from the equations (5) and (6).

<sup>4</sup> Denote the alternative crop by hat and its price by  $q$ . Then the target function and the first-order condition are  $\sum_{t=1}^n [(1+r)^{-(t-1)} [q\hat{f}(l) - cl - d(\hat{g}(l))]]$  and  $q\hat{f}'(l) - c - d'(\cdot)\hat{g}_l = 0$ , respectively. Interpretation is similar as above.

How do the socially and privately optimal solutions relate to each other? It depends on the degree of capitalization of the climate benefits in the price of the bioenergy crop. If no capitalization occurs, the sign of  $d'(\cdot)(g_t) - b'(\cdot)\mathcal{E}f'(l)$  in equation (5) is decisive. If the marginal climate benefits dominate the marginal runoff damages then the socially optimal fertilizer intensity is higher than the privately optimal one; otherwise it is lower. Under full capitalization, private fertilizer intensity is too high, because runoff damage is neglected. Depending on the degree of capitalization, the intermediate cases may entail too high or too low application rates.

The socially and privately optimal extensive margin of bioenergy crop cultivation is defined by the location where the social returns and private profits from bioenergy crop equal those of the alternative crop. Denote the social returns (private profits) from bioenergy crop by  $SW^*$  ( $\pi^0$ ) and those of the alternative crop by  $\overline{SW}^*$  and  $\overline{\pi}^0$ . Then socially and privately optimal extensive margin of bioenergy cultivation is defined by

$$k^* : SW^* = \overline{SW}^* ; \text{ and } k^0 : \pi^0 = \overline{\pi}^0 . \quad (8)$$

It is easy to demonstrate that the extensive margin of bioenergy crop production depends positively on the price of the bioenergy crop and climate benefits. Interestingly, the extensive margin depends positively also on the input price and nutrient runoff damage, because they decrease the returns of the alternative crop more than those of bioenergy crop. Finally, provided that the alternative crop can be harvested already in the first year  $t = 1$ , a higher interest rate impacts negatively on the extensive margin.

### 2.3 First-best design of policy instruments

As equations (5) – (8) reveal, the socially and privately optimal solutions differ. The difference between the optima is given by  $SW_t - \pi_t = -d'(\cdot)(g_t) + b'(\cdot)\mathcal{E}f'(l)$ . Thus, the privately optimal fertilizer application is either too low or too high depending on the

marginal social valuation of climate benefits and nutrient runoff damage. Thus, there is scope for intervention to correct for the difference. We examine here two alternative corrective mechanisms. First, the fertilizer intensity can be corrected in each location by a tax or subsidy, depending on the degree of capitalization and the relation of runoff damages to climate benefits. Second, if the capitalization rate is partial and the climate benefits dominate runoff damage, an output subsidy for bioenergy can be provided.

When a tax/subsidy,  $h$ , is levied on the fertilizer input the farmer's optimal choice becomes  $\pi_i = p^* f'(l) - c(1+h) = 0$ . If an output subsidy is provided for bioenergy, the optimality condition reads as,  $\pi_i = p^*(1+\theta)f'(l) - c = 0$ . Setting these conditions equal to the social optimality condition allows us to solve for the optimal tax/subsidy rates:

$$h^*(k) = \frac{(p^* - p)f'(l) + d'(\cdot)(g_l) - b'(\cdot)\mathcal{E}f'(l)}{c} \quad (9a)$$

$$\theta^*(k) = \frac{(p - p^*)f'(l) + b'(\cdot)\mathcal{E}f'(l) - d'(\cdot)(g_l)}{p^* f'(l)}. \quad (9b)$$

Under zero capitalization, the price difference in (9a) is zero and we have an input tax or subsidy depending on whether runoff damage or climate benefits dominate. Full capitalization in turn means that the price difference equals the marginal climate benefits and we obtain a tax for excessive fertilization. In the intermediate case, the degree of capitalization affects not only the rate of the tax/subsidy but it may also switch a subsidy to a tax and vice versa. The smaller is the capitalization, the more likely a subsidy is needed to internalize the climate benefits and vice versa.

A similar reasoning holds true for the output subsidy in (9b), but recall, an output subsidy is meaningful only under zero or partial capitalization (where climate benefits dominate). Finally, it is evident from (9a) and (9b) that the subsidy/tax schemes are location-specific, because the optimal fertilizer intensity differs across locations. Thus, differentiated policy instruments are called for in the von Thunen model (the results resemble differentiated policies under heterogeneous land quality, see Lichtenberg 2002, and Lankoski and

Ollikainen 2003). Furthermore, equations (9a) and (9b) also ensure that the extensive margin becomes optimal (for a similar proof under heterogeneous conditions, see the above references).

It is well known that differentiated policies with precise targeting may entail high transaction costs (see e.g. Vatn 2002 for general discussion). Thus, we complete this section by outlining an alternative second-best policy option for the case where climate benefits dominate nutrient runoff damage and promoting bioenergy production becomes socially optimal.

Suppose that the government pays a lump-sum area payment,  $A$ , for bioenergy production in each location. Then private profits are given by

$$\pi = \sum_{t=2}^n \left[ (1+r)^{-(t-1)} (pf(l) - cl) + A \right] - \bar{C} + A \quad (10)$$

Obviously, the area payment  $A$  does not affect optimal fertilizer intensity. However, it makes bioenergy production more profitable relative to the alternative crop, therefore affecting the extensive margin. If the aim is to expand bioenergy crop production towards the socially optimal level, the government should use an area payment that makes the private restricted profit function equal to that under the alternative exogenous crop at the socially optimal location  $k^*$ . Hence, choosing  $A = \bar{\pi}^0(k^*) - \pi^0(k^*)$  will yield the socially optimal extensive margin. However, as a result the total amount of bioenergy produced still differs from the socially optimal level of production, because the fertilizer intensity has not been corrected to the socially optimal level.

### 3. The empirical application of the von Thunen model

We illustrate our results empirically using data from the Ostrobothnia region, which is located in Western Finland. The Ostrobothnia region has two main power plants in Seinäjoki and Pietarsaari, which have a potential capacity to produce 242 GWh and 310

GWh of electricity from reed canary grass, respectively. If the average energy yield of reed canary grass is 19.17 MWh per ha (with 6 tonnes/ha yield level), then the above data would mean that 12 624 hectares and 16 171 hectares of arable land are required to be allocated to reed canary grass cultivation, respectively.

Arable land in the Ostrobothnia region is dominated by three soil types: fine sand (32.4 %), coarse sand (17.9 %), and organic (including peat) soils (21.9 %). While all soil types suit reed canary grass cultivation, the yields are highest in organic soils. Oats is the most typical cereal crop in Ostrobothnia, representing 39.6 % (37%) of total cereal cultivation area in 2004 (2005) (Yearbook of Farm Statistics 2005). Among cereals cultivated in Finland oats is best suited to peat and organic soils. Therefore, we take oats as our alternative crop.

### 3.1. Parametric version of the model

Both reed canary grass and oats are transported by road transportation (truck and trailer unit).<sup>5</sup> To keep the analysis clear, we assume that the power plant and the oats processing industry are located in the same location. Following Flyktman and Paappanen (2005) we employ the following cubic transportation cost function,  $\omega$ , for reed canary grass (transported as round bales):  $\omega^1 = \alpha k^3 - \beta k^2 + \chi k + \eta$ . For oats, the Finnish data suggests a linear transportation cost:  $\omega^2 = \varphi + \gamma k$ . Thus, the net prices of both crops are given by  $p = \hat{p} - \omega^1$ , and  $q = \hat{q} - \omega^2$ , respectively, where hat refers to the prices at the gate of the power plant ( $p$ ) and the processing firm ( $q$ ).

We use the following Mitscherlich nitrogen response function for both crops:  $y^i = \mu^i (1 - \sigma^i e^{-v^i l^i})$ , where  $y$  is yield per hectare,  $l$  is nitrogen use per hectare, and  $\mu$ ,  $\sigma$  and  $v$  are parameters. These parameters are estimated for oats by Bäckman *et al.* (1997). On the basis of Finnish experiments and nitrogen use recommendations for different soil

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<sup>5</sup> In what follows, the superscript 1 refers to reed canary grass and superscript 2 to oats ( $i = 1,2$ ) and we use subscripts to refer to time.

textural classes (clay, sandy, and organic soils) we calibrate Mitscherlich nitrogen response function for reed canary grass to reflect average yields of reed canary grass obtained with the recommended nitrogen use per ha.

Reed canary grass is a perennial crop, which is planted for a 14 year production rotation, the annual harvests starting from the third year. Fertilizer application is technologically fixed for the first two years. The private profits from reed canary grass cultivation over any location are given by

$$\pi^1 = \sum_{t=3}^n \left[ (1+r)^{-(t-1)} (p\mu(1-\sigma e^{-\nu t}) - cl - I - K + A) \right] - \bar{C}_1 - \bar{C}_2(1+r)^{-1}. \quad (11)$$

In equation (11),  $\bar{C}_1 = E + K + A$  and  $\bar{C}_2 = cl_2 + I + K + A$  comprise the establishment and some other cost items during the first two years, 1 and 2.  $E$  is the establishment costs (fuel and labour costs of primary tillage, secondary tillage, and herbicide application, as well as fertilizer, seed and herbicide costs),  $I$  denotes the variable costs of cultivation, and  $K$  refers to fixed machinery costs. Finally, we denote the annual crop area payment by  $A$ .

The profits from the alternative crop, oats are given by

$$\pi^2 = \sum_{t=1}^n \left[ (1+r)^{-(t-1)} (q\mu(1-\sigma e^{-\nu t}) - cl - I - K + A) \right]. \quad (12)$$

The climate benefits from reed canary grass are modelled as offset benefits from emissions of peat in electricity generation. The net calorific heat value is 2.45-3.94 MWh per tonne of dry matter (with the moisture content of 15-20%) for reed canary grass and 2.68 MWh per tonne for peat. Using the average heat value for reed canary grass, 3.195 MWh/ton, the amount of peat needed to produce same amount of energy is approximately 1.19 tonnes. The emission coefficient for peat is 381.6 kg CO<sub>2</sub> / MWh. Thus, one tonne of reed canary grass can reduce CO<sub>2</sub> emissions from peat by 1219 kg CO<sub>2</sub> in power production. We use the price of emission allowances as a proxy for the marginal climate

benefits. In the first trading year (2005), the average allowance price was slightly over € 20/tonne and we use this estimate. Oats does not entail any climate benefits, but for reed canary grass we have,

$$B = \sum_{t=3}^{14} (1+r)^{-(t-1)} [20 * 1.219 * f(l)]. \quad (13)$$

We describe nitrogen runoff resulting from nitrogen application by the following per hectare runoff function  $z_i = \phi_i e^{-0.7[1-0.01l_i]}$ . Parameter  $\phi_i$  calibrates this expression so that it equals the level of nitrogen runoff generated by a nitrogen application rate of 100 kilos per hectare for oats and 60 kilos per hectare for reed canary grass. We set the parameter  $\phi_i = 11$  kg N/ha for oats and  $\phi_i = 6$  kg N/ha for reed canary grass on the basis of Finnish experimental studies (Turtola and Jaakkola 1997, Partala and Turtola 1998). For the social value of nitrogen runoff damage  $d$  we use an estimate € 1.6 per reduced kg of nitrogen provided by Vehkasalo (1999).<sup>6</sup> Hence, the present value of runoff damage differs between the crops and is defined by

$$D^1 = \sum_{t=1}^2 (1+r)^{-(t-1)} (1.6 * 6 * e^{-0.7[1-0.01\bar{l}]}) + \sum_{t=3}^{14} (1+r)^{-(t-1)} (1.6 * 6 * e^{-0.7[1-0.01l]}) \quad (14)$$

$$D^2 = \sum_{t=1}^{14} (1+r)^{-(t-1)} (1.6 * 11 * e^{-0.7[1-0.01l]}). \quad (15)$$

Finally, we assume that the social returns of crops include the social benefit of retaining land in farming. We denote the annual benefit by  $\psi$ , so that the discounted benefits over the rotation period,  $\Omega$ , are

$$\Omega = \sum_{t=1}^n (1+r)^{-(t-1)} \psi \quad (16)$$

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<sup>6</sup> The nitrogen runoff function is based on Simmelsgaard (1991). Details of its calibration to Finnish conditions can be found in Lankoski and Ollikainen (2003).

By assumption, this benefit corresponds to Less Favoured Area (LFA) support in the EU, and is paid annually.

Combining equations (11) – (16), and neglecting the area payment  $A$ , allows us to express the social welfare function for reed canary grass and oats production as,

$$SW^1 = \pi^1 - D^1 + B + \Omega \quad (17)$$

$$SW^2 = \pi^2 - D^2 + \Omega \quad (18)$$

These equations will be maximized with respect to fertilizer use to provide the results in section 4.

### **3.2 Production costs, prices and support payments for reed canary grass and oats**

All prices, costs and support payments are calculated as from year 2006 (see Tables 1 and A1). We use data from production cost calculations (ProAgria 2006) for both crops in support region C2, where Ostrobothnia belongs. Table 1 shows the cost structure for reed canary grass and oats cultivation. Oats is an annual crop and its cost items are annual whereas for reed canary grass cost items are presented separately for establishment (once per rotation), management (12 times per rotation) and rotation ending (once per rotation). To make comparison more transparent the total costs of both crops are annualised present values.

We assume in Table 1 that the cereal farmer owns the basic machinery that is required to cultivate cereals and this same machinery is suitable for establishing reed canary grass as well. Thus, farmer bears the fixed costs of machinery (such as depreciation, interest, insurance) regardless of which crop is cultivated. For the estimation of capital and labour costs we use a standard activity set for field operations: primary tillage (mouldboard ploughing), seedbed tillage (harrowing), planting, and herbicide application. Field operations are conducted annually for oats and once per rotation period for reed canary grass. The capital cost is based on the machinery required for the above field operations

and machinery expense per hectare (which is measured by depreciation costs). The labour cost is based on estimate of hours/ha for different operations and the farmer's wage rate per hour. We assume that for both crops the farmer uses contractor services for harvesting (which in the case of oats covers both harvest and grain drying).

**Table 1.** *Production costs of reed canary grass and oats, €/ha (own calculations on the basis of ProAgrid 2006).*

<b>Oats, €/ha</b>		<b>Reed canary grass, €/ha</b>	
Seed	42	<b>1. Establishment (once per rotation)</b>	<b>321</b>
Herbicide	34	• Seed	70
Fuel and lubricant	31	• Fertilizer (1. and 2. year)	145
Harvest	87	• Herbicide	21
Grain drying	42	• Labour	85
Labour cost	130	<b>2. Management (12 times per rotation)</b>	<b>118</b>
Fixed costs of machinery	144	• Material costs (plastics and nets)	33
		• Harvest	80
		• Labour, fuel and lubricant cost	5
		<b>3. Costs of rotation ending (once per rotation)</b>	<b>25</b>
		• Herbicide	19
		• Labour	6
		<b>4. Fixed costs of machinery</b>	<b>144</b>
<b>Annualised total costs</b>	<b>411</b>	<b>Annualised total costs</b>	<b>218</b>

As Table 1 reveals, the annualised total cost difference between the two crops is large. The cost difference is €238 per ha in favour of reed canary grass. The difference in labour costs is especially remarkable. The rest of the parameter values, including prices and area payments, are reported in Appendix 1, Table A.1.

#### 4. Reed canary grass cultivation: optima and policies

We start by reporting the socially optimal production of reed canary grass and oats, which provide a benchmark for our policy analysis. The results are reported in terms of nitrogen application rate, production, environmental effects and social welfare (SW).

##### 4.1 The social optimum

Society chooses for each location the crop that produces the highest social welfare (SW) when environmental effects are accounted for as in equation (8). The socially optimal use of nitrogen, production, environmental effects and the overall social welfare for each location and for both crops are given in Table 2.<sup>7</sup>

**Table 2.** *Social optimum: fertiliser use, crop production and social welfare under oats and reed canary grass cultivation.*

Location, $k$	Reed canary grass					Oats			
	Nitrogen use, kg/ha	Production, kg/ha	N-runoff, kg/ha	CO <sub>2</sub> – reduction tonnes/ha	SW, €/ha	Nitrogen use, kg/ha	Production, kg/ha	N-runoff, kg/ha	SW, €/ha
0	67.4	6216	6.5	7.58	108.9	75.8	3087	9.29	7.2
10	59.8	5985	6.0	7.30	80.8	74.9	3076	9.23	2.3
20	57.8	5918	5.9	7.22	74.2	74.7	3074	9.21	1.0
30	55.9	5849	5.7	7.13	68.0	74.4	3071	9.20	-0.2
40	53.9	5777	5.6	7.04	62.2	74.2	3068	9.18	-1.5
50	51.9	5703	5.5	6.95	56.7	74.0	3065	9.17	-2.7
60	49.9	5625	5.3	6.86	51.5	73.7	3062	9.15	-3.9
70	47.9	5544	5.2	6.76	46.5	73.5	3060	9.14	-5.2
80	45.9	5460	5.1	6.66	41.8	73.3	3057	9.12	-6.4
90	43.9	5373	5.0	6.55	37.3	73.0	3054	9.11	-7.6
100	42.0	5282	4.9	6.44	33.0	72.8	3051	9.09	-8.9

<sup>7</sup> We report all results in 10 km grids in order to make reporting of results less tedious. However, results are calculated in 1 km grid.

Table 2 shows that the socially optimal nitrogen application for both crops varies and decreases with respect to the distance to the power plant (reed canary grass) or processing plant (oats). Oats has higher nitrogen intensity than reed canary grass in every location. Due to the increasing transportation costs, the nitrogen intensity decreases more steeply for reed canary grass than for oats. The CO<sub>2</sub> tonnes/ha replaced by reed canary grass decreases with decreasing yields. Still, at a distance of 100 km, reed canary grass can offset almost 6.5 tonnes of the CO<sub>2</sub> from the burning of peat. Due to lower nitrogen intensity and smaller propensity to runoff, reed canary grass has a lower per ha nitrogen runoff than oats in all locations, the difference being on average about 40%.

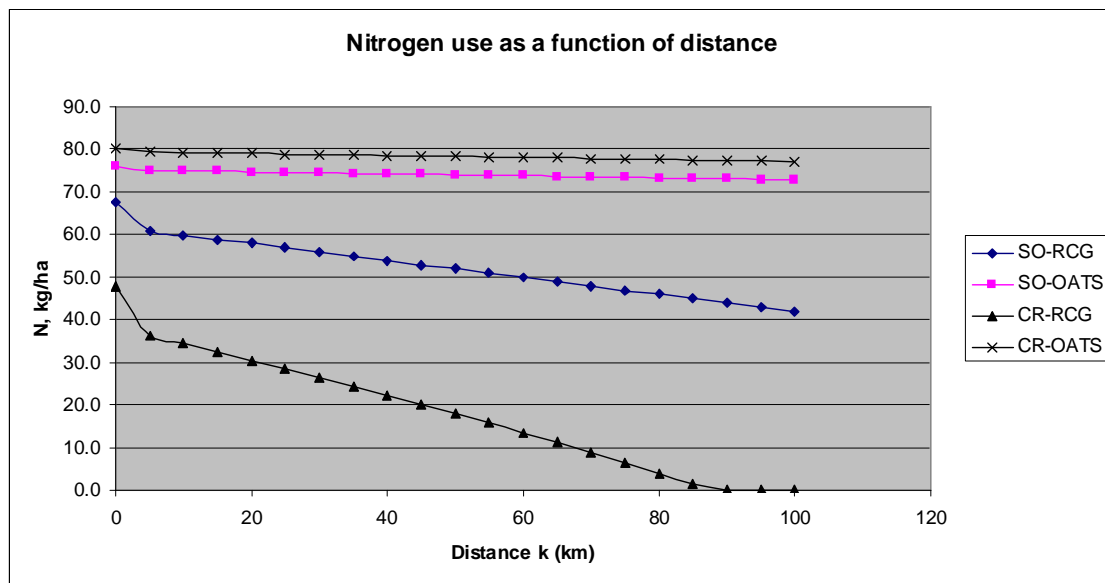
Table 2 reports the environmental effects in physical terms. How do they relate in terms of social valuation of climate benefits and runoff damages? The climate benefits are obtained by multiplying the CO<sub>2</sub> offsets in each location by the allowance price 20 €/tonne. Multiplying, the nitrogen runoff kg/ha by the damage estimate €1.6 per kg of N produces the runoff damages in each location. For example, in location 10 we find that climate benefits are €146/ha, whereas runoff damage is €9.6/ha. For all locations the climate benefits clearly dominate nitrogen runoff damage. Because of the climate benefits reed canary grass produces higher social returns than oats in every location. Oats produces positive social returns only up to location 20 and after that the returns become negative.

#### **4.2 Comparison of the social optimum and the current policy regime**

We next compare the social optimum with the current policy regime to see what kind of incentives current policy provides for bioenergy crop production. Recall that, the European Union pays a lump-sum energy crop payment (€45/ha). This area payment together with the Common Agricultural Policy and national policy instruments determine the competitiveness of reed canary grass production relative to oats. We condense our findings in Figures 1 and 2. Details are presented in Appendix 2, Table A.3, which shows e.g. the privately optimal fertilization rate under the current policy regime.

In Figure 1, RCG denotes reed canary grass, SO denotes the socially optimal solution and CR refers to current policy regime. The graphs in Figure 1 indicate the optimal nitrogen intensity as a function of distance for the social optimum and current policy.

**Figure 1.** Socially (SO) and privately optimal (CR) nitrogen use as a function of distance for reed canary grass (RCG) and oats.



From Figure 1, fertilizer intensity under current policy in oats cultivation is higher than the socially optimal application. For reed canary grass, current policy implies suboptimal and deeply decreasing nitrogen intensity with respect to distance. Therefore, from location 80 onwards the nitrogen intensity becomes so small that yield decreases below the critical yield level of 3000 kg/ha required for the energy crop payment. Thus, no energy crop payment is paid for location 80 km and thereafter (see Figure 2). The optimal nitrogen application rate becomes zero in location 90 km.

Figure 2 shows the annualized net present value profits (€/ha) for both crops in the current policy regime. Reed canary grass 1 describes the current level of support payments, whereas reed canary grass 2 assumes the same level of support payments for

both crops (the energy crop payment is €45/ha and the difference in regional payments is €5/ha and those are removed from reed canary grass).

**Figure 2.** Profits (€/ha) for oats and reed canary grass as a function of distance  $k$  under current policy regime.

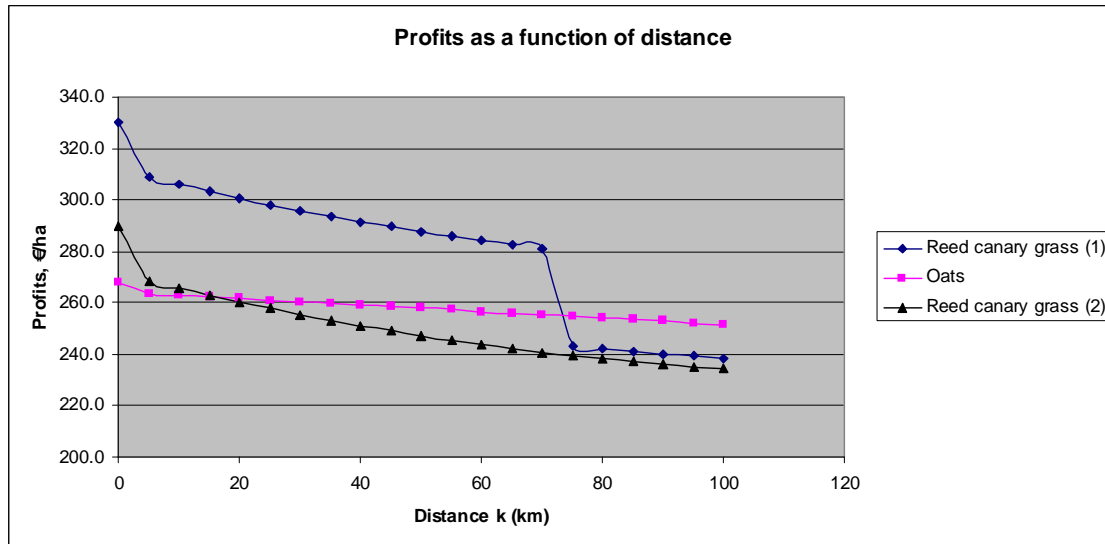


Figure 2 shows two crossing points. Under the same area payments, oats cultivation becomes more profitable from location 20 onwards, as the reed canary grass 2 line cuts oats from above. Thus, using 20 kilometres as a ray of the circle, the total profitable land area for reed canary grass would be 1256 km<sup>2</sup> (125 600 ha). In the Ostrobothnia region, the total arable land area was 243 800 hectares in year 2005 making 20% of total land area (Yearbook of Farm Statistics 2005). Consequently, using 20% as a share of arable land in the total land area, and the 20 km wide profitable distance for reed canary grass, would cover 25 120 ha of arable land, accounting for 10.3% of the total arable land in the region.

In the case of reed canary grass 1, the energy crop area payment as well as slightly higher regional area payment makes reed canary grass more profitable up to location 80. From location 80 onwards reed canary grass yield decreases below the critical yield level required for the energy crop payment and thus there is significant drop in its profitability

as shown in Figure 2. Thus, we find that under the current policy regime the profitability of reed canary grass is mainly driven by support payments (totalling €591/ha), which are €50/ha higher than the corresponding support payments for oats (totalling €541/ha).

We, finally, compare the environmental effects associated with the social optimum and the current policy regime in Table 3.

**Table 3.** *Environmental effects: social optimum (SO) and current policy regime (CR).*

Location, <i>k</i>	Reed canary grass N-runoff, kg/ha		Reed canary grass CO <sub>2</sub> reduction, tons/ha		Oats N-runoff, kg/ha	
	<i>SO</i>	<i>CR</i>	<i>SO</i>	<i>CR</i>	<i>SO</i>	<i>CR</i>
0	6.5	5.2	7.58	6.76	9.29	9.57
10	6.0	4.4	7.30	5.97	9.23	9.51
20	5.9	4.2	7.22	5.70	9.21	9.50
30	5.7	4.1	7.13	5.40	9.20	9.48
40	5.6	3.9	7.04	5.07	9.18	9.46
50	5.5	3.7	6.95	4.69	9.17	9.45
60	5.3	3.5	6.86	4.27	9.15	9.43
70	5.2	3.3	6.76	3.78	9.14	9.42
80	5.1	3.1	6.66	3.21	9.12	9.40
90	5.0	3.0	6.55	2.73	9.11	9.38
100	4.9	3.0	6.44	2.73	9.09	9.37

For reed canary grass the social optimum implies higher nutrient runoff and CO<sub>2</sub> replacement than under the current policy, because current policy fails to account for the climate benefits of reed canary grass cultivation. For oats we have the opposite situation: the current policy leads to higher nitrogen intensity and nitrogen runoff than the social optimum.<sup>8</sup>

<sup>8</sup> We did not include external costs from fuel consumption in cultivation and transportation of the crops. Oats cultivation consumes 150 liters of fuel per ha annually and cultivation of reed canary grass consumes 56 liters per ha. This strengthens the case for reed canary grass. However, oats has clear advantage as regards to fuel consumption in transportation. Using the average distance of 50 kilometres and socially optimal levels of production for both crops we estimate that for the required transportation capacity the fuel consumption is 7.7 liters per ha for reed canary grass and 1.9 liters per ha for oats. Hence, the overall difference is 88.2 liters in favor of reed canary grass.

### 4.3 Optimal policy instruments

The theoretical model established the need for government intervention to correct for negative and positive externalities related to runoff damage and climate benefits. The sign and the level of the instruments depend on the degree of capitalization of climate benefits in the price of reed canary grass. Therefore, we derive the first-best policy instruments assuming zero and partial capitalization. For the latter we use the current Finnish practice where the ETS allowance premium paid over the ordinary gate prices is on average €7/tonne at the gate<sup>9</sup>. The location-specific instruments, the nitrogen and output subsidies, are reported in Table 4 for reed canary grass under zero (the first figure) and partial capitalization (the second figure).

Table 4 confirms the theoretical finding that the optimal instrument rates differ over the various locations under both zero and partial capitalization. In line with the theoretical analysis, the optimal rates of instruments are much lower under partial than zero capitalization. As expected, the optimal nitrogen tax on oats varies little and decreases only slightly over the various locations. The optimal output subsidy rate increases in distance and shows a rather dramatic (considerable) increase over the various locations in the case of zero (partial) capitalization. While the rate of subsidy for the nitrogen input almost doubles over the distance of 100 kilometres, the output subsidy increases by a factor of nine (two), although the increase in the absolute quantity is not as big as the unit is kg.

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<sup>9</sup> Assume that the listed gate price reflects the true marginal costs of production. Then the estimate of the gate price with full capitalization of climate benefits can be developed as follows. Solve first the social optimum in the presence of climate benefits but in the absence of nutrient runoff. Define then the gate price that makes private profits (in the absence of instruments) equal to the social welfare. This produces a gate price €45.4/tonne.

**Table 4.** *First-best policy instruments under zero and partial capitalisation of climate benefits.*

Location, <i>k</i>	Reed canary grass		Oats		
	Nitrogen subsidy <i>h</i> , % and (€/kg)		Output subsidy $\theta$ , % and (€/kg)		Nitrogen tax $\tau$ , % and (€/kg)
0	<b>49</b> (0.567)	<b>44</b> (0.51)	<b>96</b> (0.0201)	<b>79</b> (0.0181)	<b>8.97</b> (0.1040)
10	<b>59</b> (0.688)	<b>41</b> (0.47)	<b>146</b> (0.0210)	<b>68</b> (0.0144)	<b>8.91</b> (0.1033)
20	<b>62</b> (0.723)	<b>41</b> (0.48)	<b>166</b> (0.0212)	<b>70</b> (0.0140)	<b>8.89</b> (0.1032)
30	<b>65</b> (0.759)	<b>45</b> (0.52)	<b>189</b> (0.0214)	<b>82</b> (0.0147)	<b>8.88</b> (0.1030)
40	<b>69</b> (0.795)	<b>46</b> (0.53)	<b>218</b> (0.0216)	<b>85</b> (0.0145)	<b>8.87</b> (0.1028)
50	<b>72</b> (0.834)	<b>50</b> (0.58)	<b>256</b> (0.0217)	<b>102</b> (0.0152)	<b>8.85</b> (0.1027)
60	<b>75</b> (0.873)	<b>52</b> (0.60)	<b>304</b> (0.0219)	<b>108</b> (0.0151)	<b>8.84</b> (0.1025)
70	<b>79</b> (0.915)	<b>54</b> (0.62)	<b>374</b> (0.0220)	<b>115</b> (0.0149)	<b>8.82</b> (0.1023)
80	<b>83</b> (0.957)	<b>55</b> (0.64)	<b>472</b> (0.0222)	<b>124</b> (0.0149)	<b>8.81</b> (0.1022)
90	<b>86</b> (0.999)	<b>58</b> (0.67)	<b>620</b> (0.0223)	<b>136</b> (0.0149)	<b>8.79</b> (0.1020)
100	<b>90</b> (1.048)	<b>64</b> (0.74)	<b>935</b> (0.0224)	<b>176</b> (0.0158)	<b>8.78</b> (0.1018)

Thus, in comparison to the input subsidy, a high level of output subsidy is required to produce the social optimum. This becomes more evident once the required government budget outlays for input subsidy and output subsidy are calculated. While the budgetary cost burden of the input subsidy for nitrogen is on average €42.6/ha (range 38.2 – 44.0), the output subsidy costs almost three times as much, on average €123.1/ha (range 118.5 – 125.5) under zero capitalization. The corresponding figures under the partial capitalization are €29.8/ha (range 27.6 – 34.4) for the input subsidy and €86.4/ha (range 80.1 – 112.3) for the output subsidy. Thus, the required government budget outlays for the input subsidy is on average only 34.7% (34.5%) of the required output subsidy under zero (partial) capitalization.

The reason for the difference in instruments lies in the way these instruments work. The input subsidy is targeted directly to nitrogen application by decreasing the effective input

price. The output subsidy works through increasing the value of the marginal product of nitrogen, which is decreasing in fertilizer use. Due to the decreasing marginal productivity of fertilizer, a change in the input price changes nitrogen application more than the corresponding output price change. Thus, under zero and partial capitalization we obtain an interesting and unconventional result: in contrast with the commonly expressed recommendation not to promote the application of fertilizers, we find that a fertilizer subsidy is a superior instrument in promoting bioenergy crop production.

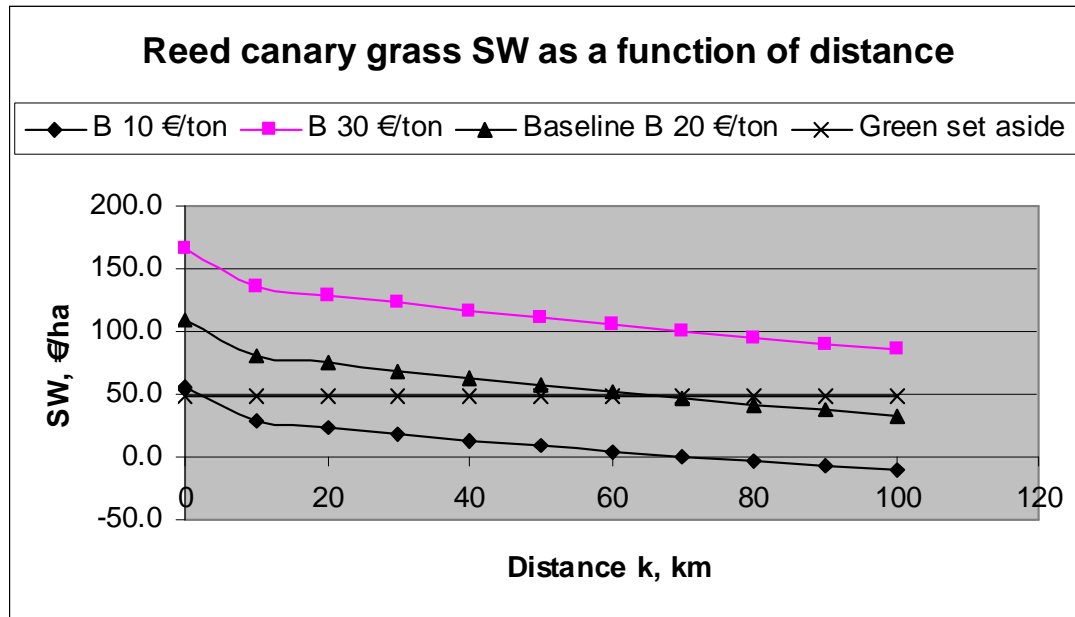
#### **4.5 Sensitivity analysis**

We performed a sensitivity analysis to examine how the key parameters affect social returns of reed canary grass cultivation. A part of this analysis was to examine how robust the findings are with respect to the chosen alternative crop. Therefore, we chose green set-aside to serve as the alternative crop for reed canary grass.

The social returns for green set aside were calculated using the following items that are the same over all locations. The social value of retaining land in agriculture is the same as before (€233/ha), the annualised cost of green set aside establishment and management are €20.5/ha, the fixed costs of machinery are €144/ha and nitrogen runoff damage from green set aside is €8/ha. This totals €48.8/ha as the present value net social return for green set aside. In comparison to baseline results (Table 2) these social returns are higher than those of oats in every location; they also exceed the returns of reed canary grass from location 70 onwards.

To see how alternative values for climate benefits (B) affect the social returns of reed canary grass, we used €10/ton and €30/ton as alternative permit prices (these values are very close to the lowest and highest values of permit prices during year 2005). The effects are shown in Figure 3 (for details see Table A2.).

**Figure 3.** Social welfare ( $SW$ , €/ha) of reed canary grass under alternative values of climate benefits ( $B$ ) in comparison to green set aside with fixed social returns.



In Figure 3, the baseline case (€20/ton) has the switching point in location 70 km for reed canary grass when green set aside is the alternative crop. For a low estimate of climate benefits (€10/ton), the extensive margin is reduced to 10 km. If climate benefits are high (€30/ton), the extensive margin shifts outwards, much beyond 100 kilometres.

We condense the rest of the sensitivity analysis in Figure 4, which represents the effects of transportation costs and nutrient runoff damage. Given that a cheaper transportation technology is under intensive investigation, we decrease the transportation costs by 20%. For the marginal nutrient runoff damage, we apply a 50% higher estimate, which is sometimes reported in Finnish studies. A higher runoff damage estimate is linked to our baseline case (€20/ton) and to a lower climate benefit estimate (€10/ton). Note that a higher runoff damage estimate also affects the social welfare performance of green set aside, so that fixed social returns are reduced to €45.6 per ha.

**Figure 4.** Social welfare (SW, €/ha) of reed canary grass under reduced transportation costs and higher nutrient runoff damage under two alternative values of climate benefits in comparison to green set aside with fixed social returns.

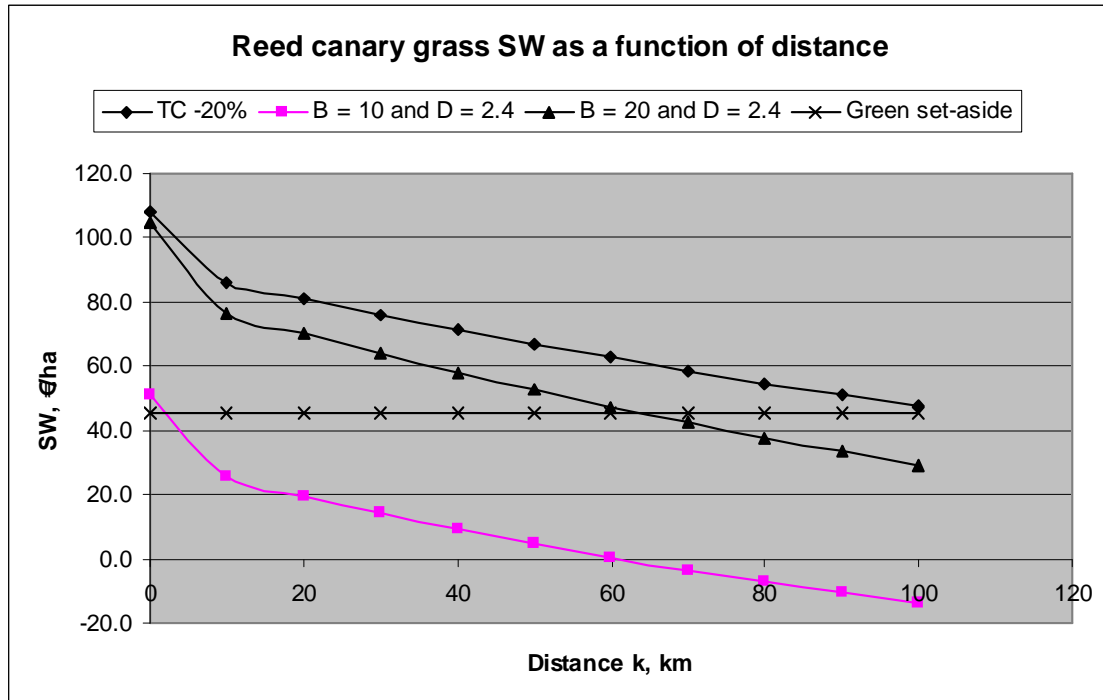


Figure 4 shows that lower transportation costs increase returns to reed canary grass, so that the switching point is moved from 70 km to slightly over 100 km. A combination of high runoff damage estimate and low climate benefit estimate make reed canary grass an inferior option relative to green set aside as from location 10 km. In comparison to the baseline, a higher runoff damage estimate does not greatly change the relative social profitability of reed canary grass and green set aside since the higher runoff damage estimate affects both land use forms.

## 5. Conclusions and policy implications

The European Union's programmes for promoting the use of renewable energy resources and the introduction of the European Union Emission Trading Scheme (EU-ETS) have increased interest in bioenergy crop cultivation. Bioenergy crops can offset fossil fuels in electricity production and thereby bring climate benefits to society. This paper examined the social returns of bioenergy crop cultivation in a von Thunen framework, when the

climate benefits are taken explicitly into account. We assumed that agricultural land is homogenous but transportation costs increase with respect to distance. Although offsetting emissions from fossil fuels in electricity production, the cultivation of bioenergy crops causes nutrient runoff to waterways.

We demonstrated that increasing transportation costs determine differing fertilizer application intensities in each location and the extensive margin of production in both the social and private optima. Provided climate benefits are only partially priced by the market, the privately optimal fertilizer application rate is lower than the socially optimal rate across all locations. Under full capitalization the privately optimal fertilizer application is higher than the socially optimal rate because of runoff damages. Thus, differentiated, location-specific policy instruments are needed – input or output subsidies in the former case, and an input tax in the latter case.

The model was applied to the cultivation of reed canary grass in Finnish agriculture, where reed canary grass offsets CO<sub>2</sub> emissions from peat in electricity production. The climate benefits were valued by the permit price in the EU-ETS. Using oats as an alternative crop, we demonstrated that cultivation of reed canary grass is socially optimal at a distance greater than 100 kilometres from the power plant and can replace more than 6.5 tonnes of CO<sub>2</sub> emissions from peat if climate benefits are valued at € 20/ton. However, it reduces to 70 kilometres if green set-aside is the alternative land-use form.

Our analysis shows that promoting energy crop cultivation is socially optimal when climate benefits dominate runoff damages and the market only partially prices climate benefits, which empirically holds for reed canary grass in Finland. An input subsidy on fertilizer application turned out to be a preferable policy instrument compared to an output subsidy, because it directly targets the privately suboptimal fertilizer application rate. Moreover, we found that current agricultural policies do not provide the best incentives to encourage bioenergy crop cultivation. Thus, a reform that promotes energy crop cultivation is needed. A good second-best choice is to increase the area payment (which is a part of CAP-policy) designed specifically to energy crops. By using an area

payment any target area for bioenergy crop production can be achieved, although fertilizer application rates will still differ from the socially optimal rates.

Our analysis relied on the assumption of competitive markets for bioenergy crop. Due to steeply increasing transportation costs the local power plants may, however, act like local monopsonists. On the basis of economic theory we know that they would equalize the marginal buyer cost with the value of the marginal product from the energy crop. This would lead to a market distortion with too low demand for bioenergy crop. It would be an interesting future research topic to examine to whether and to what extent this kind of distortion takes place. Another topic for future work is to focus closely on the social life cycle costs and benefits of using bioenergy crops to provide electricity and heat.

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## Appendix 1.

**Table A.1** *Parameter values in the numerical application.*

<i>Parameter</i>	<i>Symbol</i>	<i>Value</i>
Gate price of reed canary grass	p	€0.021/kg
Gate price of oats	q	€0.107/kg
Price of nitrogen fertilizer	c	€1.16/kg
Transportation costs:		
Reed canary grass	$\eta$	1.55
	$\chi$	0.0538
	$\beta$	0.00014
	$\alpha$	0.000000356
Oats	$\varphi$	1.5
	$\gamma$	0.05
Mitscherlich nitrogen response function	$\mu_1$	7650
	$\mu_2$	3670
	$\sigma_1$	0.7075
	$\sigma_2$	0.7075
	$v_1$	0.0197
	$v_2$	0.0197
Area payments:	A	€/ha
CAP (same for both crops)		161
LFA (same)		233
Agri-environmental payment (same)		117
Regional area payment		
Reed canary grass		35
Oats		30
Energy crop payment (only for Crop 1)		45
Nitrogen leakage at average nitrogen use	$\phi_1$	6 kg/ha
	$\phi_2$	11 kg/ha
Capital cost	K	€144/ha
Social benefit of retaining land in agriculture (same for both crops)	$\Psi$	€233/ha

**Notes:** All prices, support payments and costs are from the year 2006. The price of nitrogen is calculated on the basis of a compound NPK fertilizer.

**Table A.2.** *Sensitivity analysis of social optimum: transportation costs (TC -20%), climate benefit estimates B, nutrient runoff damage estimate D and the socially optimal nitrogen use (average in bold, range in parentheses) and extensive margin in comparison to green set aside.*

	<b>Baseline</b>	<b>B=10; D=1.6</b>	<b>B=30; D=1.6</b>	<b>B=10; D=2.4</b>	<b>B=20; D=2.4</b>	<b>TC -20%</b>
<b>Nitrogen use, kg/ha</b>	<b>56.3</b> (49.9-67.4)	<b>46.9</b> (41.6-52.3)	<b>67.6</b> (60.1-78.8)	<b>45.1</b> (39.9 - 50.4)	<b>54.7</b> (48.1-65.2)	<b>56.2</b> (48.4 -63.0)
<b>Extensive margin</b>	70	10	100	10	70	90

## Appendix 2. Reed canary grass and current policy regime in Finland

This Appendix describes the details of the current actual production of reed canary grass and oats in Finland. The private profits from cultivation were defined in equation (11) of the text. The location-specific fertilizer intensities and production per hectare are given in Table A.3.

**Table A.3.** Current policy: fertiliser use, crop production and profits under oats and reed canary grass cultivation.

Location, <i>k</i>	Reed canary grass			Oats		
	N- use, kg/ha	Production, kg/ha	Profits, €/ha	N- use, kg/ha	Production, kg/ha	Profits, €/ha
0	48.0	5547	330.0	80.2	3135	267.9
10	34.3	4895	305.9	79.2	3125	262.9
20	30.4	4675	300.6	79.0	3122	261.6
30	26.4	4430	295.8	78.8	3120	260.3
40	22.2	4157	291.5	78.5	3117	259.1
50	17.9	3849	287.6	78.3	3114	257.8
60	13.5	3501	284.1	78.0	3112	256.6
70	8.8	3100	281.1	77.8	3109	255.3
80	3.9	2635	242.1	77.6	3107	254.1
90	0.0	2238	240.0	77.3	3104	252.8
100	0.0	2238	238.3	77.1	3101	251.6

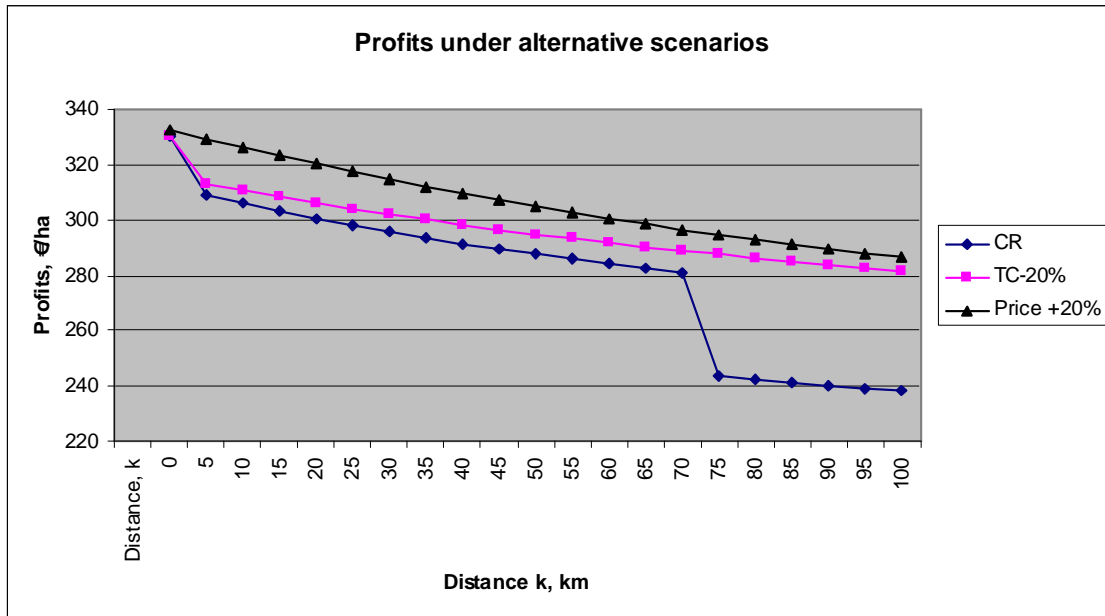
In Table A.3, the profits for reed canary grass dominate in the neighborhood of the power plant but from location 80 km onwards oats becomes more profitable. The associated nutrient runoff and climate benefits were collected in Table 3 of the text.

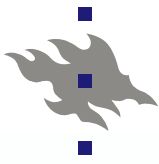
We examined how the private profitability of reed canary grass cultivation depends on exogenous parameters under the current policy regime. The results are described in Figure A.1; it shows the effects of a 20% reduction in transportation costs and a 20% increase in the price of reed canary grass paid by the power plant. Both increase the profitability of reed canary grass in comparison to the baseline. The higher output price provides higher profits than lower transportation costs in every location. However, the difference diminishes in distance as can be seen from the figure.

We also examined how the extensive margin changes if the alternative crop is green set aside instead of oats. The present value of profits from green set aside under the current agricultural support system is €185.2/ha in every location. Thus, from profits reported in

Table A.3 we find that the private profitability of green set aside is always inferior to both reed canary grass and oats under the current policy regime.

**Figure A.1.** Profits (€/ha) for reed canary grass under current policy (CR), under reduced of transportation costs (TC-20%) and increased output price (Price +20%).





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